EFFECTS FROM LOG-YARD STORMWATER RUNOFF ON THE MICROALGAE Scenedesmus subspicatus

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ABSTRACT

The effects posed by stormwater runoff from an industrial log-yard on the microalgae Scenedesmus subspicatus have been investigated. The effects of stormwater runoff sampled during two rain events on the microalgae were determined in different concentrations (% v:v). The test organisms were exposed to the stormwater runoff 96h and the percentage change in growth rates in relation to a control culture were measured at exposure times of 24h, 48h, 72 and 96h. Whereas the runoff from the first rain event posed in most cases growth stimulatory effects, the runoff from the second rain event inhibited algae growth. Differences in runoff physico-chemical characteristics combined with the hydrological factors of each rain event explained these opposite effects. The occurrence of the "first flush" phenomenon was studied and the results confirmed such occurrences in the second rain event. On the basis of normalized inhibitory effects and runoff volume, it was found that 42, 51 and 50% of the inhibitory effects during exposures of 24, 48 and 72h were associated with the initial 4% of the total discharged volume. To conclude it can be stated that since the stormwater runoff of only 2 rain events posed negative effects on the test organisms, the potential environmental threats posed by these waters during the entire hydrological year needs to be better understood in order to propose an adequate management program.

KEYWORDS

Toxicity, first flush, wood-industry, stormwater, growth inhibition

1 INTRODUCTION

It is widely known that non-point source pollution from urban stormwater runoff discharges a significant number of xenobiotics to water recipients [1], affecting aquatic organisms and deteriorating the ecosystem [2,3]. Although knowledge of stormwater runoff and pollutants transport from urban areas increase, little is known about stormwater runoff generated at industrial sites and their respective environmental impacts.

Among a number of industrial activities that might pose harmful effects to aquatic ecosystems, industrial areas that handle and store wood materials such as log-yards, sawmills and other wood-based by-products, raise concerns due to potential release of organic and inorganic compounds. According to [4], toxic effects on fishes and invertebrates are related to naturally occurring wood-related compounds such as aldehydes, phenols, terpenes and others. Furthermore, tannins, lignin; tropolones, resin acids and toxic metals such as zinc, aluminium and copper are of greatest concern due to their contribution to log-yards runoff toxicity [5,6,7].

Physico-chemical characteristics of stormwater runoff from log-yards are related to (i) the species of trees stored; (ii) the proportion of runoff that comes in contact with stored wood; (iii) log-yard size; (iv) the period the logs are stored; (v) the ratio between logs surface area to log-yard area, etc. Consequently, stormwater runoff from log-yards might have different characteristics and pose varied environmental effects as well.

The intra-storm variability of toxic effects posed by rainfall-runoff of highway areas has been previously described [8]. Those authors observed that higher toxic effects on the early stages of the runoff were gradually reduced, and that 90% of toxic effects were associated with the first 20% of the total runoff volume. Even though the toxic effects posed by wood-related compounds have been demonstrated [9,10,11,6,4], little is know concerning the intra-storm magnitude, variability and possible existence of toxic first flushes from stormwater runoff generated in log-yards.

The aim of this paper was to evaluate the magnitude and the intra-storm variability of the toxic effects posed by stormwater runoff originated from a log-yard during single storm events. Most studies previously carried out throughout entire hydrographs have focused on specific contaminant loads and less attention has been given to toxicity evaluation. Despite the useful information on levels of contamination provided, the effects on the biota are not obtained. It is essential to better understand the dynamics of adverse effects of these types of water and their magnitudes, to provide a basis for an appropriate proposal, design and implementation of stormwater runoff management strategies. For the purpose of this study the freshwater, unicellular microalga *Scenedesmus subspicatus* was selected as the test organism. Microalgae are considered ideal organisms for toxicological studies once they are primary producers and any adverse effects towards them, might influence higher levels of the food chain. Besides, microalgae have a short life cycle and are able to quickly respond to the environmental changes over a number of generations [12].

2 MATERIALS AND METHODS

2.1 Industrial site

The physico-chemical characterization of the stormwater runoff was carried out at the industrial log-yard located in Nybro, southeast of Sweden. The investigated drainage basin

had a total area of approximately $117,000 \text{ m}^2$ (Figure 1) and among different wood species stored in the area, oak (*Quercus sp.*) accounts for 40-50% [13].



Figure 1: Aerial view of the log-yard drainage basin with the drainage boundaries. Scale 1:5000 [14]

2.2 Stormwater sampling and characterization

Stormwater runoff generated by two single rain events during the summer of 2009 (June/July) were studied: (a) Rain 1, on June 18 and; (b) Rain 2, on July 1. The stormwater sampling and characterization procedures have been described by [14].

2.3 Rainfall and flow measurement

On-site precipitation was digitally recorded by an external "tipping bucket" rain gauge (SIGMA Model 2149). The standard area-velocity and Doppler principle method of flow measurement was used by a velocity submerged sensor (SIGMA Area velocity submerged sensor).

2.4 Test organisms and experimental procedures

The freshwater unicellular green alga *Scenedesmus subspicatus* was selected in this investigation. The stock culture was purchased from the algal culture collection at the University of Göttingen, Germany (SAG). The detailed description concerning the test organisms, cultivation procedures and the experimental set-up has been previously described elsewhere [14]

2.7 Data analysis

The effects of different concentrations of stormwater runoff were determined by comparing growth rates of exposed to non-exposed (negative controls) *S. subspicatus* in different exposure periods (24, 48, 72 and 96h). The average specific growth rate for *S. subspicatus* was given by:

$$\mu = \frac{\left(\ln F_t\right) - \left(\ln F_0\right)}{\Delta t} \tag{eq 1}$$

where μ is the growth rate (d⁻¹), F_t and F_0 are the biomass measured as *in vivo* fluorescence (RFU) at time *t* and t_0 respectively, and Δt is the total exposure period (*t*- t_0) in days.

On the basis of calculated growth rates, the % inhibitions were obtained as follows:

$$I\% = \frac{(\mu_c - \mu_e)}{\mu_c} \times 100 \tag{eq 2}$$

Where I% is the percentage of inhibition, μ_c and μ_e are average growth rates of negative controls and exposed *S. subspicatus* respectively. I% > 0 and I% < 0 were considered in this current work as inhibitory and stimulatory effects respectively.

Values of EC₅₀ were calculated by selecting the most approximate non-linear regression models. One-sample *t*-test was carried out to verify the significance of algal growth rate changes, regardless of being inhibition or stimulation. One-way ANOVA and post-hoc Tukey's multiple comparison test were used to assess the existence of toxic first flush. All the statistical analysis was performed by GraphPad Prism (version 5.02 for Windows, San Diego, USA) assuming Gaussian distribution and variance homogeneity in a significance level of $P \le 0.05$.

3 RESULTS AND DISCUSSION

3.1 EC 50 values

EC50 values of the July 1 rain in different exposures and runoff elapsed times are presented in Table 1. It was not possible to calculate values of $EC50_{96}$ (at the 96th hour) due to low inhibitory effects within the entire range of stormwater concentrations. The concentration-response curves with considerable small slopes indicated that regardless the exposure to different stormwater concentrations; the measured effects were nearly constant after 96-h period.

3.2 Rain event 1 – June 18, 2009

The stormwater runoff discharged on June 18, in most cases, stimulated *S. subspicatus* growth (Figure 2). Stimulatory effects might be related to the presence of some compounds in the stormwater. It has been reported that a number of trace metals are used by living organisms to stabilize protein structures, to catalyze enzymatic reactions and to facilitate electron transfer reactions in the photosynthetic and respiratory processes [15,16]. Harmless effects towards

Daphnia sp. reproduction from log-yards runoff in contact with two different wood species (*Picea abies* and *Pinus sylvestris*) have been reported [6].

According to the statistical tests (one-way ANOVA and post-hoc Tukey's multiple comparison test), considerable intra-storm variation and significant differences (p < 0.05) of the effects on *S. subspicatus* were observed after the first 48 h of exposure. Runoff samples collected on the last stages of the runoff - the last 110 minutes - posed significant higher stimulatory effects (Tukey's test; p<0.05) after 72 and 96 h exposure in comparison to the runoff samples collected throughout the first 100 minutes (Figure 2c1-c2; 2d1-d2).

Different exposure concentrations (% v:v) caused variable effects towards *S. subspicatus* and reasonably the lower the stormwater concentration (% v:v): the lower the observed stimulation or inhibitory effects (Figure 2a-d).

3.3 Rain event 2 – July 1, 2009

As shown in Figure 3, stormwater runoff of July 1 posed in most cases high inhibitory effects (one-sample *t*-test; p<0.05) to *S. subspicatus* growth reaching up to 100% in some cases (Figure 3a1). However, the inhibitory effects decreased progressively with the exposure time, and a considerable reduction in the effects was observed when comparing exposures of 24h and 96h. This might be related to an inherent process of the test organisms to detoxify and overcome the initial hazardous environment suggesting that time of exposure played an important role over the effects towards algae growth (Figure 3b-d). It was observed that the period required to overcome toxic effects was positively related to the stormwater concentrations.

In some cases it was observed a shift from inhibition to growth stimulation in which can be explained by the possible transformation and/or degradation (by e.g. light, bacteria) of certain compounds that were causing toxic effects in the first days

Runoff Elapsed time (min)	EC50 ₂₄	95% CI	8 EC504	95% CI	EC5072	95% CI	EC509 6	95% CI
15	9.5	7.3-12.2	19.3	11.4-32.6	45.9	17.8-118.7	ND	ND
40	10.0	7.5-13.5	31.0	11.3-84.9	60.6	12.6-292.4	ND	ND
65	13.2	9.6-18.5	23.5	6.8-81.0	69.2	7.0-671.0	ND	ND
90	11.9	4.6-31.0	32.5	9.8-108.0	176.1 ^a	ND ^a	ND	ND
115	21.2	13.0-34.4	106.8	ND	ND	ND	ND	ND

Table 1: EC50 values with 95% confidence intervals calculated for exposures of 24, 48, 72 and 96 h. Stormwater runoff event July 1. Values given in % v:v [14]

ND-Not determined due to low inhibitory effects; ^a imprecise theoretical model



Figure 2: S. subspicatus growth rate % in comparison to the control throughout the rain event of June 18 in exposures of (a) 24h; (b) 48 h; (c) 72h and; (d) 96h. Each row represents stormwater concentrations of (1) 50; (2) 12.5; (3) 3.13; (4) 0.79 and; (5) 0.2 % (v:v). Whiskers – Standard deviation. White bars – not significant (one-sample t-test, p > 0.05. [14]



Figure 3: S. subspicatus growth rate % in comparison to the control throughout the rain event of July 1 in exposures of (a) 24h; (b) 48 h; (c) 72h and; (d) 96h. Each row represents different stormwater concentrations of (i) 50; (ii) 12.5; (iii) 3.13; (iv) 0.79 and; (v) 0.2 % (v:v). Whiskers – Standard deviation. White bars – not significant (one-sample t-test, p > 0.05). [14]

3.4 Toxic first flush effects: normalized toxicity vs normalized runoff volume

To remove the effects of runoff flow rates on the interpretation of toxicity results, an evaluation on the basis of intra-storm normalized toxicity proportion is recommended [8]. In the current study, the normalized toxicity is expressed in terms of acute toxic units ($TU_{50}=100/EC50$).

The occurrence of toxic first flush was evident by taking into account 24; 48 and 72-h exposures, once a greater normalized toxicity proportion in the earlier stages of the runoff event in comparison to the normalized discharged volume were observed (*see Figure 4*). It can be observed that 42, 51 and 50% of the toxic effects during exposures of 24, 48 and 72h respectively were associated with the first 4% of the total discharged volume and 13% of the total storm duration (Figure 4).



Figure 4: Discharged volumes and toxic units (TU50) as a function of storm elapsed time on a normalized basis for acute exposures of 24, 48 and 72 h. [14]

3.5 Rain events characteristics

The differences observed in the magnitude and nature of *S. subspicatus* responses – from growth stimulation to inhibition – when comparing both rain events, might have been influenced by the hydrological factors (rain intensity, precipitation heights, antecedent dry period and duration) and the stormwater runoff physico-chemical characteristics. Possible relationships between *S. subspicatus* responses, storm characteristics and chemical pollutant concentrations were evaluated based on the hydrographs and pollutographs (COD, pH and conductivity).

Significant differences (*t*-test, p<0.05) observed between metals concentrations, pH and conductivity of the two runoff events, might explain the opposite responses of the test organisms. The results showed also that the pH of the runoff from July 1 was significantly lower (p<0.05) suggesting that to some extent, low pH might have been responsible to the observed *S. subspicatus* growth inhibition Increased toxicity with

lower pH's might be related to an increase of metals solubility and bioavailability of prevailing free metal ions [17].

Differences in metal concentrations were also observed and the concentrations of Cr, Ni, V, Ba, and Zn were significantly higher (*t*-test; p<0.05) in the rain event of July 1, with the exception of Cu (p<0.05). The toxic effects due to metal contents in wood-related areas are reported in the literature by [5,6]. Furthermore, according to [18] metabolic processes such as photosynthesis have been shown to be very sensitive to the presence of toxic metals.

Other aspects that could have contributed to the responses from *S. subspicatus* are the hydrological factors involved in the rainfall-runoff process such as storm duration, precipitation heights and average rainfall intensity. Whereas 3mm of precipitation was recorded during 115 minutes in June 18 (average intensity of 1.56 mm/h), only 1.25 mm was recorded during 45 minutes in July 1 (average intensity of 1.66 mm/h). Lower amounts of stormwater, but a nearly similar rainfall average intensity during the event of July 1 might be the cause of higher concentrations of metals and H⁺ detected within the runoff water. Assuming that similar rainfall intensities as observed, are able to wash-off the log-yard the same mass of contaminants; higher concentrations will be detected in lower amounts of water, posing then higher toxic effects.

Similar results concerning the roles played by hydrological aspects over the toxic effects of runoff have been reported by [2]. The authors also observed that at the same rain intensities, rainfall duration and degrees of toxic effects were inversely related.

4 CONCLUSIONS

The runoff originated from two different rain events over the same log-yard area posed different effects on the freshwater, green microalgae *S. subspicatus*. Whereas the runoff of the first rain event had no negative effects to *S. subspicatus* posing in most cases stimulatory effects, the runoff from the second rain event inhibited algae growth. The differences in runoff physico-chemical characteristics combined with the hydrological factors (precipitation and duration) of each rain event explained the opposite effects on *S. subspicatus* observed during the current investigation.

Despite the lack of negative effects to *S. subspicatus* growth, stimulatory effects posed by the first runoff event indicate that the necessity of treating these waters before discharging into natural recipients can not be ruled out since the stormwater runoff might contribute to environmental impacts such as algae blooming and eutrophication.

The intra-storm magnitude and variability of the inhibitory effects posed by the 2^{nd} runoff event confirmed the hypothesis of toxic first flush phenomenon. The evaluation on a normalized basis showed that 42, 51 and 50% of the inhibitory effects during exposures of 24, 48 and 72h are associated with the initial 4% of the total discharged volume and 13% of the storm duration.

To conclude, the current investigation brings a better understanding of a subject that has not been widely explored: the intra-storm magnitude and variability of the potential environmental effects posed by log-yards runoff. Toxicity studies can provide useful information for an appropriate design and implementation of stormwater runoff management strategies. The fact that negative effects were observed in the two runoff events analyzed in this study, raises concern about the potential environmental threats posed by runoff originated from wood-based industrial areas during the entire hydrological year.

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